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**Geogenic factors as drivers of microbial community diversity in soils overlying  
polymetallic deposits**

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5 **GEOGENIC FACTORS AS DRIVERS OF MICROBIAL COMMUNITY**  
6 **DIVERSITY IN SOILS OVERLYING POLYMETALLIC DEPOSITS**  
7

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42 **Abstract**

43 This study shows that the geogenic factors landform, lithology and underlying mineral  
44 deposits (expressed by elevated metal concentrations in overlying soils) are key-drivers  
45 of microbial community diversity in naturally metal-rich Australian soils with different  
46 landuse, *i.e.*, agriculture vs. natural bushland. 168 soil samples were obtained from two  
47 metal-rich provinces in Australia, *i.e.*, the Fifield Au-Pt-field (New South Wales) and the  
48 Hillside Cu-Au-U-rare-earth-element (REE) deposit (South Australia). Soils were  
49 analyzed using three-domain multiplex terminal-restriction-fragment-length-polymorphism  
50 (M-TRFLP) and PhyloChip microarrays. Geogenic factors were determined using field-  
51 mapping techniques and analyses of >50 geochemical parameters. At Fifield, microbial  
52 communities differed significantly with geogenic factors and equally with landuse  
53 ( $P<0.05$ ). At Hillside, communities in surface soil (0.03-0.2 m depth) differed significantly  
54 with landform and landuse ( $P<0.05$ ). Communities in deeper soils (>0.2 m) differed  
55 significantly with lithology and the mineral deposit ( $P<0.05$ ). Across both sites, elevated  
56 metal contents in soils overlying mineral deposits were selective for a range of bacterial  
57 taxa, most importantly *Acidobacteria*, *Bacilli*, and *Beta-* and *Epsilon-Proteobacteria*. In  
58 conclusion, long-term geogenic factors can be equally important in determining soil  
59 microbial community diversity than landuse.

60

61 **INTRODUCTION**

62 Determining the drivers of microbial community composition in soils is challenging,  
63 because soils are complex ecosystems containing numerous ecological niches. This  
64 results in an immense diversity with up to thousands of taxa per gram of soil (1).  
65 Landuse, climate, vegetation, soil-type and anthropogenic pollution are strongly linked to  
66 soil microbial community structures, functions and activities at many sites (2-4). In  
67 particular, agricultural practises are known to drive differences in soil microbial  
68 communities (5). Influences of geogenic factors, e.g., landform, underlying lithologies and  
69 mineral deposits on soil microbial communities have received little attention. However,  
70 three studies of soils from Switzerland and Nepal have shown that  
71 geological/mineralogical factors can significantly affect species assemblages and  
72 functions (6-8).

73 A primary goal of geomicrobiological research is to link microbial communities and  
74 metal cycling in metallogenic environments, and to determine how communities are  
75 structured due to metal-associated drivers (9). Microorganisms play a pivotal role in the  
76 biogeochemical cycling of many metals, particularly those essential for cell function, e.g.,  
77 Co, Cu, Fe, Mg, Mn, Mo, Na, Ni, V, W and Zn, and those oxidized or reduced in catabolic  
78 reactions to gain metabolic energy, e.g., As, Fe, Mn, Mo, Sb, Se, Sn, Te, U and V (10).  
79 Therefore, metal-rich environments are useful model systems allowing links between  
80 microbial taxa and geochemical parameters to be identified (11).

81 To date, mine tailing- and acid mine drainage sites have been primary *foci* for  
82 assessing community compositions and functions at metal-rich sites (12-15). In contrast,  
83 few studies have assessed microbial communities in naturally metal-rich soils with low to  
84 moderate metal contents. These sites are highly abundant and typical examples include

85 soils overlying/surrounding buried mineral deposits, where physical and (bio)geochemical  
86 cycling has led to the formation of metal enrichment zones (16). A recent Canadian study  
87 has shown that distinct microbial community assemblages were present in the glacial  
88 cover overlying a buried volcanogenic massive sulfide (VMS) deposit (17). In this study, a  
89 strong correlation between Zn and Cu concentrations, total biomass and abundances of  
90 methanotrophic bacteria was observed (17). Using culture-based approaches,  
91 correlations between abundances of *Bacillus cereus* spores and the presence of Au and  
92 its pathfinder elements (*i.e.*, As, Ag, Bi, Cu, Mo, Se, Te) in soils overlying Au deposits in  
93 Belgium, the United States and Australia have been observed (18-20). In Western  
94 Australian soils overlying Cu-Pb-Zn-deposits the solubilization, transport and deposition  
95 of metals is mediated by resident plant- and microbial communities (21, 22).  
96 Subsequently, elevated concentrations of mobile metals in the soils are related to  
97 changes in the microbial community composition. In particular, highly mobile elements,  
98 *e.g.*, S, Zn, Cl and Al, were implicated as drivers of bacterial community structures across  
99 these sites (21, 22). A study of 187 soils collected at four naturally auriferous (*i.e.*, Au-  
100 containing) areas in remote Australia has shown that microbial communities and  
101 functional potentials differ significantly with landform, soil depth, lithology and Au-deposits  
102 (23). This demonstrated that geogenic factors are important drivers of microbial  
103 community diversity at sites where anthropogenic landuse is minimal and uniform across  
104 sampling localities. However, as geogenic influences are likely to develop over extended  
105 'geological' time periods, their effects may be masked by short-term changes in landuse  
106 (24, 25). This has been illustrated in a study assessing the impact of short-term changes  
107 of landuse, *e.g.*, by over-sowing with exotic grasses and legumes, on soil microbial  
108 community structures and functional potentials at four tussock-based grassland sites in

109 New Zealand (26). Results have shown that soil bacterial and fungal communities and  
110 functional capabilities were strongly influenced by landuse, but unaffected by sampling  
111 locality, *i.e.*, geographic distance and environmental setting.

112 We hypothesize that influences of geogenic factors (*i.e.*, landform, underlying  
113 lithology and mineral deposits) on soil microbial community assemblages are equally  
114 important than those of anthropogenic landuse. Therefore, the aims of this study are to i)  
115 assess if differences in landuse mask influences of geogenic factors, and ii) determine  
116 the dominant bacterial taxa differentiating soils overlying metal deposits from adjacent  
117 background soils. To achieve this, samples were collected from two metallogenic sites in  
118 Australia. The Fifield Pt-Au-field site is Australia's largest platiniferous regions hosting  
119 range of Alaskan-type primary Pt-, VMS- and hydrothermal Au deposits. The Hillside  
120 deposit is located in the Gawler Craton, which also hosts one of the world's largest Cu-  
121 Au-U deposits, Olympic Dam. Soil microbial communities were characterized using M-  
122 TRFLP and high-density phylogenetic microarrays (PhyloChip G2), field-  
123 mapping/geochemical analyses were used to determine geogenic factors and multivariate  
124 statistical approaches were used to test the hypotheses.

125

## 126 **MATERIALS AND METHODS**

### 127 **Field sites and sampling**

128 Soils with elevated natural metal contents and adjacent background soils were  
129 collected from two Australian sampling areas, *i.e.*, Hillside and Fifield (Fig. S1). The  
130 Fifield Au- and Pt-field is situated approximately 380 km north-west of Sydney, Australia  
131 at 32°50'33.48" S 147°28'5.38" E (Fig. S1). It was the largest Australian producer of Pt in  
132 the 1900s (27). The climate at Fifield is semi-arid with most plant growth occurring in

133 summer and temperature limiting growth in winter. Soils are largely residual, with little  
134 material being transported into the area (28). They are classified as Red Sodosols  
135 following the Australian soil classification scheme (28). Landscape evolution in the area  
136 commenced in the Early to Middle Devonian (29). Subsequent periods of weathering,  
137 laterization and fluvial erosion have occurred since (29). This has resulted in an  
138 undulating landscape with distinct erosional, colluvial, alluvial and depositional landforms  
139 (Australian landscape classification system; 30). At Fifield, 75 soil samples were collected  
140 across a 20 km transect covering Au, Pt and base metal deposits with differing underlying  
141 lithologies (Ordovician and Silurian-Devonian (meta)-sediments and metal-rich  
142 intrusions), landforms (erosional, colluvial, alluvial and depositional), and differing  
143 landuse (grassland for cattle and natural *Eucalypt* bushland) in May 2011 from the A-  
144 horizon (at 0.1-0.15 m depth). Note: Due to the extended weathering history and long  
145 periods of tectonic stability of the Australian continent, landscape evolution and  
146 weathering have occurred for millions to hundreds of millions of years leaving behind  
147 deeply weathered (down to 1000 m) *in situ* or transported weathered materials, with  
148 landscapes looking to the “untrained” eye flat with little or no relief (31). Therefore a  
149 specific classification system was developed to classify these landscapes in a  
150 geomorphological context; this system was used for this study (30).

151 The Hillside site is located close to the township of Ardrossan in South Australia, at  
152 34°32'04.32" S 137°52'41.81" E (Fig. S1). The site is situated in the Gawler Craton, which  
153 also hosts the Olympic Dam Fe-oxide Cu-Au-U-REE deposit. The highly weathered *in*  
154 *situ* lithology is covered by well-sorted, rounded, Aeolian sediments consisting of  
155 spherical quartzose, calcareous sands, nodular- and hardpan carbonates (32). The  
156 climate at Hillside is semi-arid, with average yearly minimum of 10.6°C and maximum of



157 22.6°C. Much of the Hillside area has been cleared for agriculture, especially wheat  
158 production. Prior to clearing, the vegetation consisted mostly of mallee woodland  
159 dominated by red mallee (*Eucalyptus socialis*); remnants of this vegetation still occur  
160 throughout landscape. The primary metal deposit occurs in Proterozoic basement rocks  
161 and ore metals are hosted by the sulfides bornite and chalcopyrite; in the oxidized zone  
162 overlying the sulfidic zone secondary Cu carbonates and U-minerals have formed (32).  
163 The contemporary landscape consists of subdued relief and is lower at the coast and  
164 higher inland (32). 93 soil samples were collected across four transects in April 2011.  
165 Transects covered areas of differing landuse (wheat cropping and native Mallee  
166 woodland), landforms (erosional, colluvial and depositional), geophysical responses  
167 (airborne electromagnetic indicative of different lithologies) and depths. Surface soils  
168 were collected from 0.03 to 0.2 m depths. Deeper soils were collected directly above the  
169 carbonate hardpan, below depths of 0.2 to 0.5 m.

170 At each sampling site, six 50 mL centrifuge tubes of soil were collected under field-  
171 sterile conditions. Samples for DNA-extraction were frozen on-site. Tubes stored at  
172 ambient temperature were used for geochemical analyses.

173

#### 174 **Geochemical characterization**

175 After homogenization, Fifield soil samples were microwave digested in  
176 concentrated *aqua regia*. Concentrations of major metals were determined by inductively  
177 coupled plasma-optical emission spectrometry (ICP-OES; Spectro ARCOS SOP,  
178 Germany); minor- and trace metals were determined by inductively coupled plasma-mass  
179 spectrometry (ICP-MS; Agilent 7500ce, USA; following 23). Total C and N were  
180 determined by high-temperature combustion (Formacs analyser; Skalar Inc., USA);

181 electrical conductivity (E.C.) and pH were measured in 1:5 soil to water extracts. For  
182 Hillside samples, an existing dataset of soil geochemical parameters was used (32).  
183 Geochemical parameters (data available on request) were categorized into seven groups,  
184 including solution parameters (E.C. and pH), and six elemental groups based on a  
185 modified Goldschmidt element classification system (33, 34): (i) Pathfinder- (Fifield: Ag,  
186 As, Au, Bi, Mo, Pb, Pd, Pt, Se and W; Hillside: As, Ag, Au, Ba, Bi, Cu, Mo, Pb, Sb, Se,  
187 Te, Th, U, V and W), (ii) biophile- ( $C_{tot}$ ,  $N_{tot}$ , P and S), (iii) rare earth (Be, Ce, Dy, Er, Eu,  
188 Gd, Ho, La, Lu, Nb, Nd, Pr, Sm, Tb, Tm, Y and Yb), (iv) chalcophile- (Cd, Cu, Ga, Ge, Sn,  
189 Ti, Tl and Zn), (v) lithophile- (Al, Ca, Cr, Cs, Hf, K, Li, Mg, Na, Nb, Rb, Sc, Sr, Th, U, and  
190 V), and (vi) siderophile- (Fe, Co, Mn, Ni, Te) elements.

191

## 192 **Assessment of community assemblages and functional potential**

### 193 ***Nucleic acid extraction, quantification and quality control***

194 For M-TRFLP, DNA was extracted in duplicate from 0.25 g of homogenized field-  
195 fresh soils using the PowerSoil DNA Isolation kit (MoBio, USA) with a mechanical  
196 disruption step of  $5 \text{ m s}^{-1}$  for 20 s on a Bio101 FastPrep bead-beater. Duplicate extracts  
197 were pooled and used for further analyses. For microarray analyses, DNA was extracted  
198 in duplicate from 10 g of soil using the PowerMAX Soil Mega Prep DNA Isolation kit  
199 (MoBio, USA). DNA quality was assessed spectrophotometrically (NanoDrop ND-1000,  
200 USA): Only DNA with 260 : 280 and 260 : 230 ratios of 1.8 and 1.5, respectively, was  
201 pooled prior to further analyses. The total amount of DNA extracted was quantified using  
202 Quant-it Picogreen dsDNA reagent (Invitrogen, USA) on a MX3000P qPCR (Stratagene,  
203 USA); unknown concentrations were compared against a standard curve derived from  
204 known concentrations of  $\lambda$ -phage DNA.

205

206 **Multiplex terminal restriction fragment length polymorphism (M-TRFLP)**

207 Bacterial, archaeal and fungal communities were characterized using M-TRFLP of  
208 the small subunit rRNA gene (35). Multiplex-PCR (25  $\mu$ L volume) used Qiagen HotStar  
209 *Taq* chemistry and thermocycling consisted of 30 cycles of 95°C for 30 s, 55°C for 30 s,  
210 72°C for 60 s and a final extension step for 10 min at 72°C (35). PCR products were  
211 purified (Wizard<sup>®</sup>SV; Promega), and 100 ng digested with 20 U of *MspI*, *HaeIII* and *TaqI*  
212 for 3 h at 37°C or 65°C. Capillary separation of TRF's was conducted at the Australian  
213 Genome Research Facility and TRF's scored using GeneMarker software (SoftGenetics,  
214 USA) at a detection limit of 200 fluorescent units; TRF's differing  $\pm$  0.5 base-pairs (bp)  
215 were binned together (23). Relative peak heights were used as measure of abundance.  
216 Richness was based on the number of unique TRF lengths obtained.

217

218 **PhyloChip analysis**

219 The PhyloChip G2 microarray was used to further characterize the bacterial and  
220 archaeal community compositions (36). Key-samples were chosen based on  
221 environmental factors, *i.e.*, landuse, soil depth, landform, underlying mineral deposits and  
222 lithology. DNA from three replicates with matching factor combinations was pooled and  
223 analyzed on one array. Note: as soils had been extracted in duplicate and pooled, each  
224 array was run with a mixed sample from six individual DNA extraction (*i.e.*, three factor  
225 replicates and two extraction repeats). Amplification of 16S rRNA genes, purification of  
226 products, labelling of DNA and hybridisation were conducted following Brodie *et al.*  
227 (2006) using the reaction chemistry described in Wakelin *et al.* (2012b; 22, 36).  
228 Hybridized arrays were stained and washed on an Affymetrix fluidic station (36). After

229 scanning, data were processed following the method outlined by Brodie *et al.* (2006) and  
230 DeSantis *et al.* (2007; 36, 37). Data were imported into PhyloTrac for scoring of taxa (38).  
231 Operational taxonomic units (OTUs) were deemed detected by a positive fraction (PF) of  
232 probe-pair matches  $\geq 0.9$ .

233

### 234 **Statistical analyses**

235 Multivariate analyses were conducted in the PRIMER software package with the  
236 PERMANOVA add-on using statistical approaches described in previously (23, 39, 40).  
237 Soil geochemical data (except pH) were log-transformed to remove skew. Data were  
238 normalized and similarity matrices based on Euclidean distances were calculated. For  
239 bacterial, fungal and archaeal community analyses, each TRF was treated as an OTU  
240 and the peak height was inferred as representing the relative abundance of that OTU.  
241 Similarity matrices were generated on square-root transformed abundance data using the  
242 Bray-Curtis method (41). For PhyloChip data (log values), a similarity matrix was created  
243 using the Gower method (42). The taxonomical hierarchy for each taxon was determined  
244 and the distribution of phyla/classes plotted; taxa representing  $<1\%$  of the total  
245 abundance were combined as 'other'. Differences in overall abundances of phyla/classes  
246 between soils from different landforms, landuse, lithologies and underlying mineral  
247 deposits, were calculated. Significance levels were determined using student's t-tests  
248 ( $P < 0.05$ ). PERMANOVA (permutational multivariate ANOVA; 40) for TRFs was used to  
249 test if between group variation (*i.e.*, location, soil depth, landuse, landform, lithology and  
250 mineral deposits) can explain a significant proportion of the total system variation (*i.e.*, if  
251 natural groupings can be detected). Balanced PERMANOVA analyses were conducted  
252 using partial sums of squares on 9999 permutations of residuals under a reduced model.

253 CAP analysis (canonical analysis of principal coordinates; 40), was used to determine if  
254 principal coordinates axes could be found that separate *a priori* defined treatments (*i.e.*,  
255 CAP analysis attempted to 'seek-out' pre-defined groups within the data cloud). Vector  
256 overlays, based on Pearson correlations, were used to explore relationships between  
257 significant individual variables and the ordination axes. CAP analyses were conducted  
258 based on the respective resemblance matrices; the significance of test effects was  
259 determined against null distributions based on 999 or 9999 permutations (random  
260 allocations) of samples (40). Distance-based linear modelling (DISTLM) was used to  
261 assess the geochemical parameters best explaining the variability within the microbial  
262 dataset (23). SIMPER (similarity percentage) analysis was used to identify  
263 taxa/classes/phyla that discriminate between locations, landforms, landuse and  
264 underlying lithology and mineral deposits. For phylum-/class-level SIMPER analysis, data  
265 were normalized to take into account differences in probe numbers between the different  
266 phyla represented on the array. Given the high number of PhyloChip variables (OTUs),  
267 interpretation of treatment effects on bacterial communities were conducted at the  
268 phylum/class level. UPGMA and maximum-likelihood (with 1000 boot strap replicates)  
269 phylogenetic trees based on PhyloChip probes for 50 OTUs that best discriminate soils  
270 overlying mineral deposits from background soils as well as taxa detected only in soils  
271 overlying the mineral deposits were constructed using GENEIOUS v. 7.0.

272

## 273 **RESULTS**

### 274 **Linking community profiles, geochemistry and environmental factors**

275 Significant links between bacterial, fungal and archaeal community assemblages,  
276 landuse, landform, underlying lithology and mineral deposits, expressed through the

277 geochemical properties of soils, were present at both locations. CAP and cluster analyses  
278 of microarray data showed that bacterial and archaeal community assemblages varied  
279 significantly between sites (Hillside v. Fifield;  $P=0.0003$ ), depths (Hillside;  $P=0.01$ ), across  
280 different landuses ( $P=0.003$ ) and with underlying mineral deposits ( $P=0.009$ ; Figs. 1 and  
281 2A; Table 1); a strong interactive effect of landform and mineral deposits was observed  
282 was also observed (Table 1).

283 At Fifield, soil geochemical properties varied significantly with lithology ( $\sqrt{CV}=4.1$ ;  
284  $P<0.001$ ), landform ( $\sqrt{CV}=3.2$ ;  $P<0.001$ ) and mineral deposits ( $\sqrt{CV}=5.3$ ;  $P<0.001$ ); no  
285 significant differences between the grazing- and bushland, were detectable (Table 2).  
286 Bacterial community assemblages (based on M-TRLFP) varied significantly with landuse,  
287 landform, lithology and mineral deposits; lithology ( $\sqrt{CV}=14.1$ ;  $P=0.01$ ) and landuse  
288 ( $\sqrt{CV}=13.9$ ;  $P<0.02$ ) were the primary discriminators (Table 2). Fungal communities  
289 varied significantly with landuse ( $\sqrt{CV}=14.9$ ;  $P=0.01$ ) and underlying lithology ( $\sqrt{CV}=16.7$ ;  
290  $P<0.01$ ), but not landform or mineral deposits (Table 2). Archaeal communities varied  
291 significantly with all geogenic factors, but not landuse (Table 2). No significant interactive  
292 effects were observed. All groups of geochemical parameters showed significant  
293 relationships with the microbial data, the strongest being the solution parameters as well  
294 as biophile- and siderophile elements (Fig. 3). Levels of Au/Pt pathfinder and REE in  
295 samples explained 14.4% and 27.3% of variation in the bacterial community composition,  
296 respectively (Fig. 3). Gold/Pt pathfinder elements explained 12.8% and 18.0% of variation  
297 in fungal and archaeal communities, respectively (Fig. 3). Solution parameters, biophile,  
298 siderophile and pathfinder elements also explained most of the variation in community  
299 assemblages detected with high density microarrays ( $P<0.05$ ; Fig. 2B).

300 At Hillside, microbial communities varied with most with soil depth; e.g., the  
301 bacterial community displayed a  $\sqrt{CV}=16.5$  ( $P<0.01$ ; Table 3) across top-soils (A-horizon;  
302 0.03-0.2 m) and sub-surface soils (B-horizon;  $>0.2$  m). No influences of either landuse or  
303 geogenic factors were detected when surface- and deeper soils were analyzed  
304 collectively (Table 3). Analyzed independently, bacterial and fungal communities in top-  
305 soil varied significantly with landuse ( $\sqrt{CV}=12.5$ ;  $\sqrt{CV}=4.7$ ;  $P<0.001$ , respectively) and  
306 landform ( $\sqrt{CV}=9.4$ ;  $\sqrt{CV}=3.4$ ;  $P<0.05$ , respectively). Archaeal communities were not  
307 linked to any of the factors tested (Table 3). In sub-surface soils, bacterial communities  
308 varied most significantly with lithology ( $\sqrt{CV}=9.2$ ;  $P=0.05$ ) and the mineral deposit  
309 ( $\sqrt{CV}=14.2$ ;  $P<0.001$ ), but not landuse or landform (Table 3). Fungal communities varied  
310 with lithology and the mineral deposit as well as landuse (Table 3). An interactive effect  
311 between landform and mineral deposit was observed for bacterial and fungal  
312 communities. Across all Hillside soils, geochemical properties varied with soil depth. The  
313 landform was a significant influence in deep soils. Pathfinder elements and solution  
314 parameters were capable of explaining 70.4% and 20.6% ( $P<0.05$ ) of variation in the  
315 bacterial community, respectively; this was confirmed by analyses linking geochemical  
316 properties and microarray data (Fig. 2C). Lithophile major elements were capable of  
317 explaining 24.2% of in fungal community; other associations were not significant.

318

### 319 **Linking taxa, geochemistry and environmental factors**

320 PhyloChip G2 microarrays were used to compare bacterial and archaeal  
321 communities representative of location, landform, landuse, lithology and mineral deposits.  
322 Across all samples, 45 phyla, 90 classes, 173 orders and 306 families were detected  
323 (Fig. 1; Table S1). Between 879 and 1879 individual taxa were observed (Fig. 1; Table

324 S1). Bacterial communities were dominated by *Proteobacteria* (*Alpha*-, *Beta*-, *Gamma*,  
325 *Delta* and *Epsilon*- subdivisions; 35.5-45.7%), *Firmicutes* (15.1-22.6%), *Actinobacteria*  
326 (12.2–17.1%), *Acidobacteria* (3.8-6.3%), *Bacteriodes* (2.6-5.7%), *Chlorofexi* (1.8-3.3%),  
327 and *Cyanobacteria* (2.1–4.0%) (Fig. 1). *Sphingobacteria*, *Verrucomicrobiae*,  
328 *Anaerolineae*, *Planctomycetacia*, *Catabacter*, *Spirochaetes*, and *Archaea* represented  
329 between 0.1 and 2.0% to the composition of prokaryotic communities (Fig. 1).

330 Prokaryotic communities at Hillside were richer in taxa (1660±101) compared to  
331 Fifield (1007±194; Table S1). All phyla/classes displayed higher numbers of taxa in  
332 Hillside compared to Fifield soils. Of these, *Alpha*-, *Beta*- and *Gamma-Proteobacteria*,  
333 *Actinobacteria*, *Bacteroides*, *Sphingobacteria* and *Spirochaetes* were the most  
334 discriminatory taxa ( $P < 0.05$ ; Table S1). At Hillside the total number of taxa in different  
335 phyla/classes did not differ between soils depth, landuse and underlying mineral deposit.  
336 However, 208 taxa were only detected in soils overlying the mineral deposit (Table S1).

337 At Fifield, significantly higher numbers of taxa ( $P < 0.05$ ) were observed in samples  
338 overlying mineral deposits (1153±170 taxa) compared to background soils (881±72 taxa;  
339 with 463 taxa occurring only in soils overlying mineral deposits (Table S1). In particular,  
340 *Alpha*-, *Beta*-, *Delta*- and *Epsilon-Proteobacteria* as well as *Acidobacteria*,  
341 *Sphingobacteria* and *Verrucobacteria* were significantly more abundant in soils overlying  
342 mineral deposits. Twenty-eight taxa occurred in soils overlying the mineral deposits at  
343 Fifield as well as Hillside; of these ten belonged to *Bacilli* (Fig. S2).

344 At Hillside higher numbers of *Beta-Proteobacteria* and *Bacilli* were associated with  
345 colluvial compared to erosional landforms (Table S1). At Fifield overall abundances  
346 between erosional and colluvial landforms were similar, alluvial sites contained higher  
347 number of *Alpha*-, *Beta*-, *Gamma*- and *Epsilon-Proteobacteria*, *Firmicutes* (*Bacilli* and



348 *Clostridia*) and *Cyanobacteria* (Table S1). In soils from cattle grassing sites significantly  
349 more *Bacilli* were detected compared to native bushland soils (Table S1). Whereas  
350 *Acidobacteria*, *Beta-* and *Epsilon-Proteobacteria*, *Cyanobacteria*, *Spirochaetes* and *Bacilli*  
351 correlate well with underlying mineral deposits, *Actinobacteria*, *Delta-Proteobacteria*, and  
352 *Bacteroidetes* correlated most strongly with background soils (Fig. 2A). SIMPER analyses  
353 on individual taxa also showed that of the 19 and 20 of the 50 most discriminating taxa  
354 between soils overlying mineral deposits and background sites were *Acidobacteria* and  
355 *Proteobacteria*, respectively (Fig. 4). *Gamma-Proteobacteria* correlated with colluvial  
356 landforms. *Actinobacteria* were closely linked to erosional landforms (Fig. 2A; Table S1).  
357 At alluvial sites a larger abundance and diversity of *Cyanobacteria* was observed (Fig.  
358 2A; Table S1).

359

## 360 DISCUSSION

361 We show that landform, underlying lithology and mineral deposits are closely  
362 related to microbial community assemblages in geologically 'older', naturally metal-rich  
363 soils. At our study sites geogenic factors are as important in explaining the variation in  
364 community assemblages as anthropogenic landuse. This, we hypothesize, is a likely  
365 result of the extended history of weathering and metal cycling at these sites. Over  
366 extended periods of weathering, metal-bearing minerals are decomposed through the  
367 interaction of biogenic and abiogenic factors (43). As a result heavy metals (e.g., Au, Hg,  
368 Pb, Ag, Cd, Cu, Zn, Ni and U) are mobilized and become bioavailable in soils (43). For  
369 example, Fe- and S-oxidizing microorganisms alter metal sulfides leading to the  
370 production of acid and the mobilization of heavy metals (44). Re-precipitation and  
371 biomineralization of metals leads to the formation of metal enrichment zones, which in-

372 turn affect microbial communities (16, 21-23). Elevated concentrations of mobile metals  
373 can select for community assemblages that are better able to deal with metal toxicity and  
374 are therefore better suited to survive in these environments (45, 46). This is often  
375 expressed as an increase in the diversity and/or abundance of metal-resistant  
376 populations (47-49) and may not be easily masked by differences/changes in landuse, as  
377 observed in soils from New Zealand, Europe and North America evolving from  
378 unweathered rock since the extensive Upper Pleistocene glaciations (*i.e.*, geologically  
379 “young” soils; *e.g.*, 26, 50-52).

380         The data presented here supports our hypothesis that longer *in-situ* weathering  
381 periods lead to larger differences in community diversity driven by geogenic factors. At  
382 Hillside, the highly-weathered Palaeoproterozoic lithology and associated metal deposit  
383 are covered by Aeolian (*i.e.*, <10,000 years old) sediments, consisting of quartzose, and  
384 pedogenic carbonates. During this time, decomposition of Cu-, Au- and U-bearing  
385 bornite, pyrite and chalcopyrite, combined with the biogeochemical cycling of these metals  
386 and their pathfinders, has led to the formation of secondary metal enrichment zones,  
387 which strongly influenced community assemblages. In total, 208 bacterial taxa (mostly  
388 *Bacilli*, *Acidobacteria*, *Alpha-* and *Gamma-Proteobacteria*) occurred only in soils overlying  
389 the Cu-Au-U-REE deposit. Assessment of overall taxa abundances at the phylum/class  
390 levels showed little differences between different factors tested (Table S1). This suggests  
391 that individual taxa were replaced at sites overlying the deposit, but that the concentration  
392 of toxic, mobile heavy metals in combination with “geologically shorter” exposure time  
393 was not sufficient to alter community composition at this level. At Fifield soils have  
394 evolved continuously from *in situ* materials for millions of years (29), leading to soils  
395 highly enriched in mobile metals. Microbial communities at this site have been subjected

396 to elevated metal concentrations for very long periods of time (potentially millions of  
397 years). Here all of the measured elemental groups were significantly linked to microbial  
398 community assemblages, with 462 taxa detected only in soils overlying the mineral  
399 deposits. Significant differences on class/phylum level were observed with *Alpha-*, *Beta-*,  
400 *Delta-* and *Epsilon-Proteobacteria* as well as *Acidobacterial* and *Cyanobacterial* taxa,  
401 which were more numerous and abundant in the metal-rich soils.

402 To further test this hypothesis and identify key-drivers affecting community  
403 assemblages, a range of studies and experiments can be conducted, including: i)  
404 determining the soil meta-genome of wide range (continental-scale) of soils with  
405 established histories of landuse, age, lithology, landform and mineral deposits; this will be  
406 important to assess if correlations observed in our study occur in general in Australia and  
407 internationally; ii) establishment of soil micro- and mesocosm experiments with well-  
408 defined model communities and *in-situ* communities from geologically “young” and “old”  
409 soils; these can be incubated with a range/combination of mobile metal ions and effects  
410 on community composition and function can be assessed; iii) in field trials diversity-  
411 disturbance responses of soil microbial communities can be measured after amendment  
412 of soils with increasing metal doses; here geologically “young” and “old” soils with  
413 different landuse can be tested to be assess, if communities in older soils react differently  
414 to those from younger environments.

415 While communities at both sites were strongly affected by location, particular  
416 groups of bacteria contributed strongly to the differences between metal-rich and  
417 background soils across both sites (Figs. 2A and 4; Table S1). The  
418 abundance/occurrence of *Acidobacteria*, *Verrucomicrobia*, *Bacilli* and *Proteobacteria*  
419 contributed most strongly to the differences between metal-rich and background soils

420 across both sites. These taxa are known for their ability to withstand elevated  
421 concentrations of mobile heavy metals and/or affect the speciation and mineralogical  
422 association of these metals as discussed below. Few *Acidobacteria* and *Verrucomicrobia*  
423 have been cultured to date, yet sequence data shows that both phyla are ubiquitous and  
424 highly abundant in many soils (53-55). A recent study completed the genomes of three  
425 *Acidobacterial* strains (56). Based on the combination of physiological and genomic  
426 evidence, the authors suggested that *Acidobacteria* are long-lived, divide slowly, exhibit  
427 slow metabolic rates under low-nutrient conditions, and are well equipped to tolerate  
428 fluctuations in hydration and high contents of mobile metals observed at both study sites  
429 (56). Hence, they are well adapted to survive at both study sites. In particular,  
430 *Acidobacteria* were commonly identified in soils and sediments overlying or containing U  
431 deposits, respectively, as well as in waste rock and mill tailings from U mines (36, 54, 56).  
432 In our study, many of the key-organisms differentiating metal-rich from background sites  
433 matched probes of an *Acidobacteria* first identified by Geissler and Selenska-Pobell  
434 (2005) from a U waste piles near Johanngeorgenstadt (Germany; 57). Other taxa were  
435 first identified in samples obtained from Au mines. *Verrucomicrobia* have also been linked  
436 to U mining environments (58), with the taxa identified here belonging to the ammonia-  
437 oxidising group. The ubiquity and abundance of *Acidobacteria* and *Verrucomicrobia* in  
438 metal-rich soils at the study sites, combined with their ability to survive in metal-polluted  
439 extreme environments, suggest that they serve functions important to biogeochemical  
440 metal cycling at the study sites, similar to those observed other U-rich sites (59).

441 *Bacilli* are halotolerant, alkaliphile/alkalitolerant and capable of forming  
442 endospores; these can persist in a dormant state and tolerate extreme conditions (60,61).  
443 In a previous studies, elevated number of *B. cereus* spores were also detected in

444 auriferous soils from Belgium, China, Argentina and Mexico, and the use of *B. cereus* as  
445 a bioindicator for Au-exploration has been proposed (18, 19). A significant increase in the  
446 number of *B. cereus* spores in auriferous soils from Tomakin was observed (20). In  
447 microcosm experiments, the abundance of *Bacilli* were also significantly higher in Au-  
448 amended compared to unamended soils (23). These data suggest that *Bacilli* are  
449 particularly well adapted to persist in commonly dry semi-arid soils displaying elevated  
450 concentrations of heavy metals.

451       Active *Bacilli* from semi-arid Australian environments may also have another  
452 capability affecting metal cycling. Research has shown that the formation of metal-  
453 anomalous pedogenic carbonates is biomediated through the activity of resident *Bacilli*,  
454 and is not simply the result of passive nucleation on inactive cells or evapotranspirative  
455 processes as previously thought (e.g., 62). Enrichment cultures from South Australian  
456 pedogenic carbonates from an area adjacent to Hillside with similar environmental  
457 conditions, consisted of *Bacillus* and related *Paenibacillus* and *Lysinibacillus* taxa (63).  
458 These cultures were shown to induce Ca-carbonatogenesis as well as the co-  
459 precipitation and subsequent enrichment of Au, U and Cu in pedogenic carbonates (63,  
460 64). This suggested that *Bacilli* contribute strongly to the formation of metal anomalous  
461 zones in carbonate rich soils, which in turn effects community composition.

462       *Proteobacteria*, especially *Alpha-Proteobacteria*, were the dominant phylum across  
463 the study sites and are also often dominant class of bacteria in metal-contaminated  
464 Australian soils (21-23). The *Proteobacteria* contain well-characterized metallophilic  
465 bacterial genera, e.g., *Cupriavidus* and *Pseudomonas*, which have also been detected in  
466 the metal-rich soils at the Fifield and Hillside sites (65-66). In addition to surviving under  
467 high metal conditions, some of these microorganisms have been shown to play a

468 functional role in the bioprecipitation and biomineralization of metals, e.g., Cu and Au (67,  
469 68). The presence of these organisms at the study sites suggests that microbial  
470 communities may be well adapted to high contents of mobile metals at the sites.

471 In conclusion, the results of this study show that geogenic factors, *i.e.*, landform,  
472 lithology and mineral deposits and associated geochemical parameters are can strongly  
473 affect the composition of microbial community assemblages in naturally metal-rich soils.  
474 The study expands on the results of earlier works (21-23, 69) in a number of important  
475 ways: (i) landuse at the study sites was not uniform and/or minimal, instead intensely  
476 agriculturally utilized soils were compared to native bushland soils; (ii) soils overlying  
477 polymetallic Cu-Au-U-REE and Au-Pt deposits were assessed, compared to earlier  
478 studies that featured soils overlying economic Au deposits; and (iii) soils from sites with  
479 strongly differing histories of landscape evolution were assessed, *i.e.*, soils resulting from  
480 *in situ* weathering (Fifield) vs. soils formed from Aeolian materials overlying heavily  
481 weathered terrain (Hillside). This indicates that geogenic factors are important for the  
482 selection of microbial community assemblages in 'geologically old' soils and landscapes,  
483 and hence may contribute to variation in the soil microbial community diversity at a far  
484 wider range of sites than previously suggested.

485

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694 **Figure legends**

695 **Figure 1** Distribution of dominant prokaryotic phyla/classes (number of OTUs in this  
696 group given in bars) and cluster analyses of community data based on  
697 individual taxa; note: classes are shown for *Proteobacteria* and *Firmicutes*;  
698 phyla with <1 % coverage are aggregate into “other”.

699 **Figure 2** Ordination plots of the first two Cap produced by CAP of Phylochip data  
700 analysed for differences in community assemblages in relation to landform  
701 and underlying mineral deposits. Vectors of Pearson’s correlations of  
702 classes/phyla (A) and geochemical parameter (B,C) overlain.

703 **Figure 3** Percentage of bacterial, fungal and archaeal soil community diversity  
704 explained by geochemical parameters; results of distance based linear  
705 regression modelling (DISTLM), based on M-TRLFP data v. geochemical  
706 parameters of Fifield samples with significance level ( $P < 0.05$ )

707 **Figure 4** 16s rRNA gene maximum-likelihood phylogenetic tree (1000 boot strap  
708 replicates) of the 50 taxa (PhyloChip) that best discriminate soils overlying  
709 mineral deposits from background soils at both sites; taxa were identified by  
710 SIMPER analyses. Note: One representative (probe-targeted) sequence per  
711 taxum was used; the tree does not represent sequences the from field sites,  
712 but is a close approximation based on probe matches; sequences were  
713 obtained from GenBank.

714



715 **TABLES**716 **Table 1**

717 Summary of CAP of the factors location,  
718 depth, landuse, lithology and mineral  
719 deposits on microbial community  
720 assemblages at the Fifield and Hillside  
721 sites; significant  $P < 0.05$  (bold font).  
722

Factor	CAP	
	Trace <sup>c</sup>	P <sub>perm</sub>
<b>Location</b>	<b>0.97</b>	<b>0.0003</b>
<b>Depth</b>	<b>0.91</b>	<b>0.01</b>
<b>Landuse</b>	<b>0.63</b>	<b>0.003</b>
<b>Mineral deposit</b>	<b>0.95</b>	<b>0.009</b>
Landform (Lf)	0.31	0.9
<b>M x Lf</b>	<b>2.82</b>	<b>0.05</b>

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724

725 **Table 2** Summary of PERMANOVA  
 726 testing of the influence of the factors landuse,  
 727 lithology, landform and minerals deposits on  
 728 microbial community assemblages (M-TRLFP  
 729 data) and geochemical parameters at the  
 730 Fifield site; significant  $P < 0.05$  (bold font).  
 731

PERMANOVA		
	( $\sqrt{\text{CV}}$ ) <sup>a</sup>	P <sub>perm</sub>
<b>Bacteria</b>		
Landuse	<b>13.9</b>	<b>0.02</b>
Lithology	<b>14.1</b>	<b>0.01</b>
Landform	<b>8.0</b>	<b>0.05</b>
Mineral	<b>6.1</b>	<b>0.01</b>
Residual	27.3	-
<b>Fungi</b>		
Landuse	<b>14.9</b>	<b>0.01</b>
Lithology	<b>16.4</b>	<b>0.004</b>
Landform	4.1	0.36
Mineral	4.8	0.16
Residual	40.6	-
<b>Archaea</b>		
Landuse	13.4	0.1
Lithology	16.7	0.08
Landform	<b>22.6</b>	<b>0.004</b>
Mineral	<b>12.5</b>	<b>0.01</b>
Residual	50.8	-
<b>Geochemistr</b>		
Landuse	1.1	0.23
Lithology	<b>4.1</b>	<b>0.0003</b>
Landform	<b>3.2</b>	<b>0.0004</b>
Mineral	<b>5.3</b>	<b>0.0001</b>
Residual	4.4	-

732 <sup>a</sup> ( $\sqrt{\text{CV}}$ ) is the square root of the component of variation, which is a  
 733 dataset dependent measure of the effect of size in units of the  
 734 community dissimilarities (*i.e.*, increasing positive values); negative  
 735 values indicate zero components (40).

736 **Table 3**  
 737 Summary of PERMANOVA testing of the influence of  
 738 the factors depth, landuse, lithology, landform and the  
 739 underlying Cu-Au-U deposit on microbial community  
 740 assemblages and geochemical parameters at the  
 741 Hillside study site; significant  $P < 0.05$  (bold font).  
 742

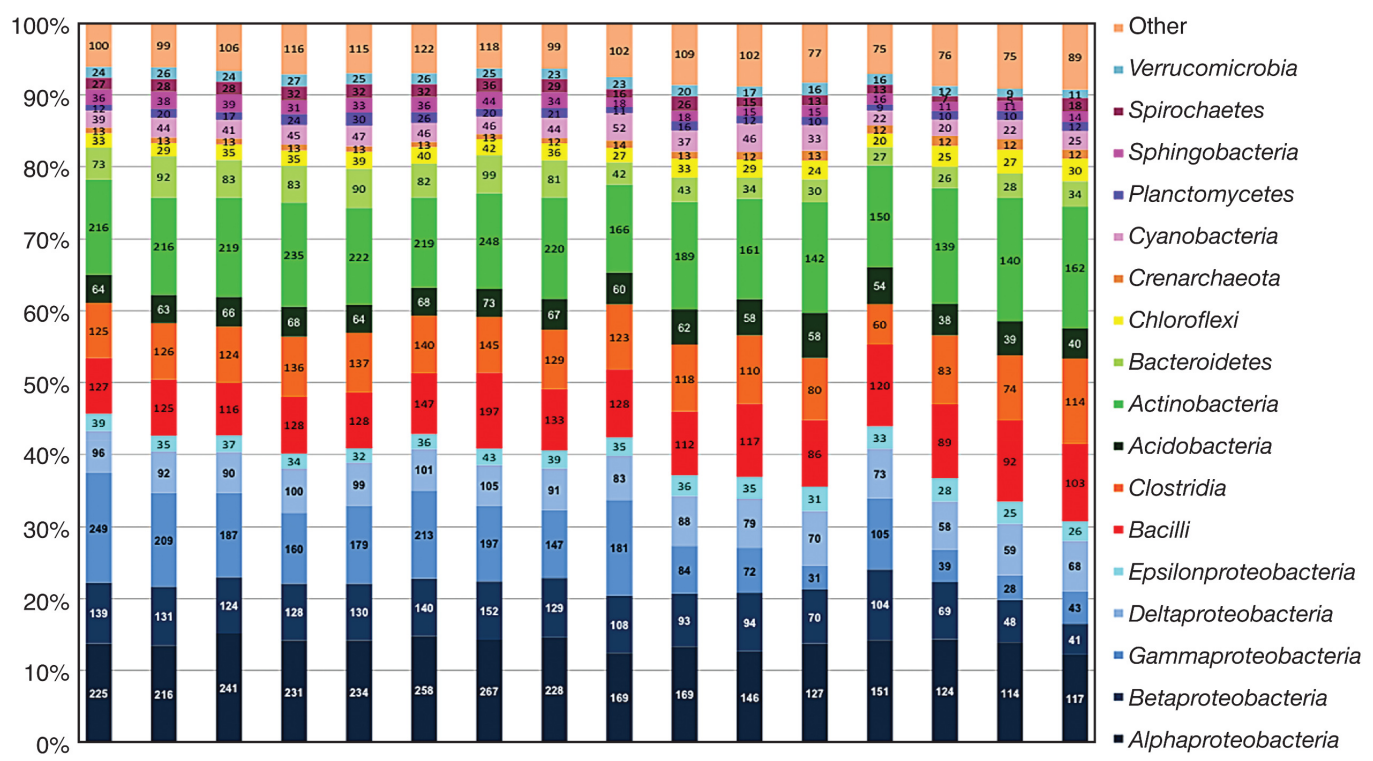
	<i>Bacteria</i>		<i>Fungi</i>	
	<b>PERMANOVA</b>	$P_{perm}$	<b>PERMANOVA</b>	$P_{perm}$
	( $\sqrt{\text{CV}}$ ) <sup>a</sup>		( $\sqrt{\text{CV}}$	
<b><i>All samples</i></b>				
Depth	<b>16.5</b>	<b>0.01</b>	<b>20.3</b>	<b>0.0006</b>
Landuse	9.6	0.3	11.1	0.2
Landform, lithology, mineral deposit	-9.0	0.9	-1.7	0.5
Residual	33.5	-	28.9	-
<b><i>Surface samples</i></b>				
Lithology	2.9	0.4	5.2	0.27
Landform	<b>9.4</b>	<b>0.04</b>	<b>15.1</b>	<b>0.02</b>
Mineral deposit	-5.8	0.8	-5.4	0.56
Landuse	<b>12.5</b>	<b>0.001</b>	16.6	0.0001
Residual	15.4	-	22.0-	-
<b><i>Deep samples</i></b>				
Lithology	<b>9.2</b>	<b>0.05</b>	18.7	0.08
Landform (Lf)	6.9	0.16	10.9	0.18
Mineral deposit (M)	<b>14.2</b>	<b>0.006</b>	<b>9.6</b>	<b>0.04</b>
Landuse	6.1	0.17	7.5	0.1
Lf x M	7.7	0.25	<b>20.0</b>	<b>0.05</b>
Residual	37.7	-	28.7-	-
	<i>Archaea</i>		<i>Geochemistry</i>	
	<b>PERMANOVA</b>	$P_{perm}$	<b>PERMANOVA</b>	$P_{perm}$
	( $\sqrt{\text{CV}}$		( $\sqrt{\text{CV}}$	
<b><i>All samples</i></b>				
Depth	<b>17.4</b>	<b>0.0001</b>	-	-
Landuse	4.8	0.35	-	-
Landform, lithology, mineral deposit	7.6	0.21	-	-
Residual	29.2	-	-	-
<b><i>Surface samples</i></b>				
Lithology	-5.4	0.54	-	-
Landform	7.8	0.24	-	-
Mineral deposit	-7.5	0.66	-	-
Landuse	14.8	0.031	-	-
Residual	23.5	-	-	-
<b><i>Deep samples<sup>b</sup></i></b>				
Lithology	7.0	0.30	<b>3.12</b>	<b>0.001</b>
Landform (Lf)	13.9	0.07	<b>2.51</b>	<b>0.05</b>
Mineral deposit (M)	6.4	0.3	-4.43	0.4
Landuse	4.2	0.4	n.t. <sup>c</sup>	n.t.
Lf x M	<b>30.0</b>	<b>0.04</b>	<b>2.51</b>	<b>0.02</b>
Residual	30.9	-	7.2	-

<sup>a</sup> ( $\sqrt{\text{CV}}$ ) is the square root of the component of variation, which is a dataset dependent measure of the effect of size in units of the community dissimilarities (i.e., increasing positive values); negative values indicate zero components (40).

<sup>b</sup> a company dataset was used for analyses, which only provided data for deep samples.

<sup>c</sup> not tested, as all geochemical analyses were derived from wheat cropping sites

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Hs	Hs	Hs	Hs	Hs	Hs	Hs	Hs	Ff	Ff	Ff	Ff	Ff	Ff	Ff	Ff	Location
su	de	su	de	de	de	su	su	su	su	su	su	su	su	su	su	Depth
no	no	no	no	yes	yes	yes	yes	yes	yes	yes	yes	yes	no	no	no	Mineral deposit
wc	wc	nb	nb	wc	nb	wc	wc	cg	cg	cg	nb	cg	nb	cg	nb	Landuse
co	co	er	er	co	er	co	er	al	er	dep	co	co	er	er	co	Landform
low	low	low	low	hi	hi	hi	hi	SD	SD	SD	SD	VMS	SD	SD	Or	Lithology

